

Phenotypic and genotypic study of antibiotic-resistant *Escherichia coli* isolates from a wastewater treatment plant in Zulia state, Venezuela.

Elba Guerrero¹, Lizeth Caraballo¹, Howard Takiff¹, Dana García² and Marynes Montiel^{3,4}

¹Laboratorio de Genética Molecular, Centro de Microbiología y Biología Celular, Instituto Venezolano de Investigaciones Científicas (IVIC), Caracas, Venezuela.

²Centro de Investigación del Agua, Universidad del Zulia, Maracaibo, Venezuela.

³Facultad Experimental de Ciencias. Universidad del Zulia, Maracaibo, Venezuela.

⁴Escuela Superior Politécnica del Litoral, Facultad de Ciencias de la Vida, Guayaquil, Ecuador.

Keywords: antibiotic resistance; *E. coli*; wastewater; phylogroups.

Abstract. Antibiotic-resistance in bacteria is a global health problem, and wastewater treatment plants can play a role in their dissemination. In this work, we used PCR and plasmid transformation to characterize antibiotic-resistance and the phylogenetic groups of *Escherichia coli* isolated from a treatment plant in Zulia, a state in western Venezuela. Thirty-six bacteria isolates were analyzed, of which 27 resulted resistant by disc diffusion primarily to tetracycline and sulfisoxazole but also to trimethoprim, chloramphenicol, ampicillin, and ciprofloxacin. The *tetA*, *sul2*, *floR*, and *blaTEM* resistance genes were frequently present and, in most cases, transferable. *dfrA12*, *tetB*, *sul3*, *sul1*, and *aacA2* genes also were detected. The integrase gene *intI1* was common in multidrug-resistant isolates. These results suggest that *E. coli* from the treatment plant is a reservoir of antibiotic-resistance genes, which signify a potential health threat. Additionally, the phylogroup C was predominant, which is unusual and may represent an adaptation of this group to environmental conditions or perhaps the most frequent phylogroup entering from the influent.

Estudio fenotípico y genotípico de aislados de *Escherichia coli* resistentes a antibióticos de una planta de tratamiento de aguas residuales del estado Zulia, Venezuela.

Invest Clin 2023; 64 (3): 296 – 307

Palabras clave: genes de resistencia; *E. coli*; aguas residuales; filogrupos.

Resumen. La resistencia bacteriana a antibióticos es un problema de salud global y las plantas de tratamiento pueden jugar un papel en su diseminación. En este trabajo caracterizamos, mediante PCR y transformación de plásmidos, la resistencia a antibióticos y los grupos filogenéticos de *Escherichia coli* aislada de una planta de tratamiento en el estado Zulia, Venezuela. Se analizaron 36 aislados bacterianos, de los cuales 27 resultaron resistentes por difusión en disco principalmente a tetraciclina y sulfisoxazol, pero también a trimetoprim, cloranfenicol y ampicilina. Los genes *tetA*, *sul2*, *floR* y *blaTEM* se encontraron comúnmente en los aislados resistentes y fueron en la mayoría de los casos transferibles; adicionalmente se detectaron los genes *dfrA12*, *tetB*, *sul3*, *sul1* y *aadA2*. El gen de integrasa *intI1* se detectó en la mayoría de los aislados multi-resistentes. Estos resultados sugieren que *E. coli* en la planta de tratamiento es un reservorio de genes de resistencia a antibióticos, lo que significa una amenaza potencial para la salud. Adicionalmente predominó el filogrupo C, lo que es inusual y podría deberse a una adaptación de este a las condiciones ambientales o podría ser el mayoritario en el influente.

Received: 28-09-2022 Accepted: 23-03-2023

INTRODUCTION

The indiscriminate use of antibiotics in human and animal medicine, as well as for prophylaxis and growth promotion in animal husbandry, threatens to reduce the effectiveness of these fundamental drugs. Antibiotic-resistant bacteria, although occurring naturally, are also released into the environment, where they may outcompete sensitive bacteria due to the presence of antibiotics and other chemical contaminants that are also released into the environment. The resistance genes in these bacteria can then be transferred to other pathogenic and non-pathogenic bacteria, thereby increasing the environmental reservoir of resistant bacteria and genetic resistance determinants.

Escherichia coli is a commensal bacterium that inhabits the intestines of humans and other animals and is often used to indicate environmental fecal contamination. Some *E. coli* are pathogens that can cause urinary tract, gastrointestinal or nosocomial infections. Antibiotic-resistance in environmental *E. coli* has also been proposed as an indicator to monitor the extent of antibiotic resistance in the environment¹.

The bacterial load of wastewater discharged into natural water bodies is significantly reduced by treatment plants. However, these plants may also promote the spread of antibiotic-resistant bacteria and resistance genes by providing favorable conditions for increasing the relative abundance of resistant bacteria and the horizontal transfer of

the genes conferring this resistance ². Although there are few treatment plants in Venezuela and much of the wastewater is discharged directly into the environment, there is a wastewater treatment plant in the state capital Maracaibo, located in the “El Tablazo” Petrochemical Complex of the Miranda municipality of Zulia state. The plant was designed so that the petrochemical industry could reuse some of its effluent water while the rest would be discharged into the giant Lake Maracaibo. There have been very few studies on antibiotic-resistance in bacteria isolated from raw or treated wastewater in Venezuela, but such studies represent essential surveillance measures to assess the extent of antibiotic-resistance in the environment and plan corrective strategies. Accordingly, we set out to perform a phenotypic and molecular study of antibiotic resistance in *E. coli* isolates from the wastewater treatment plant mentioned above.

MATERIALS AND METHODS

E. coli was isolated from water samples of the “El Tablazo” treatment plant (Miranda Municipality, Zulia State) collected from May to October 2012 for a microbiological quality evaluation. The system includes a pre-treatment to remove solids followed by absorption, biological oxidation, and a first chlorine injection. The water is then transported to the plant at “El Tablazo” and subjected to physical and biological treatment based on reactors where dissolved organic matter is removed, followed by a secondary settling. Then a first effluent is discharged into Lake Maracaibo. Another portion of the water to be used by the petrochemical complex is treated with a flocculant and chlorine.

Sampling sites were four different sections of the treatment plant: pre-treated influent (Site 1); after physical processing (Site 2); after biological processing (effluent to the Maracaibo lake, Site 3); and the chlorine disinfection point (Site 4). There

were six water samples from Site 1 and Site 4, four from Site 2, and five from Site 3.

The water samples were collected in sterile bottles and processed according to the procedures described in the Standard Methods for examination of Water and Wastewater to determine coliform by the fermentation technique ³. Samples showing growth in EC broth were streaked onto EMB agar to select typical *E. coli* colonies, which were sub-cultured in nutrient agar tubes for transport. Re-isolation was performed on McConkey agar, and colonies were cultured in LB broth and then stored in 20% glycerol at -80°C. All assays were performed on the bacteria regrown from the frozen stocks.

Biochemical identification and antibiotic susceptibility testing

Bacterial isolates were first identified with the following biochemical tests: TSI, indole-motility, methyl red, Voges Proskauer, citrate, and urea.

Resistance was assessed with the Kirby Bauer disc diffusion method, using commercial discs with the following antibiotics: tetracycline 30 µg (TE), ampicillin 10 µg (AMP), ampicillin-sulbactam 10/10 µg (SAM), sulfisoxazole 250 µg (SF), chloramphenicol 30 µg (C), trimethoprim 5 µg (W), trimethoprim-sulfamethoxazole 1.25 µg /23.75 µg (SXT), ciprofloxacin 5 µg (CIP), aztreonam 30 µg (ATM) and imipenem 10 µg. *E. coli* ATCC 25922 was used as an antibiotic-susceptible control strain. The results were interpreted according to CLSI guidelines ⁴.

PCR amplification

PCR was used to confirm the bacteria as *E. coli*, determine phylogroups, and detect the presence of resistance genes and the integrase gene *int11*. All PCR reactions were performed on boiled bacterial lysates, using Taq DNA polymerase with ThermoPol buffer (NEB), following the manufacturer's instructions, using previously reported specific primers, some of which were modified

as indicated below. PCR was performed to amplify genes conferring resistance to tetracycline (*tetA* and *tetB*)⁵, sulfisoxazole (*sul1*, *sul2*, and *sul3*)⁶⁻⁸, chloramphenicol (*floR* and *cat*)⁶, ampicillin (*blaTEM*)⁹ and trimethoprim (*dfrA12* and *dfrA7&17*)^{10, 11}. The detection of the *intI1* integrase was with primers described by Moura¹². In contrast while the phylogroup identification was performed using the quadruplex plus group C specific PCR described by Clermont *et al.*¹³ Negative controls without template DNA were included in each PCR assay. A subset of the PCR products were confirmed by DNA sequencing (Macrogen, Korea) and used as positive controls for the detection of resistance genes and *intI1*, as well as the determination of phylogroups.

The reactions were performed with previously reported primers and conditions or with the following variations: *Sul1-R* 5'-TGATCTAACCCCTCGGTCTCT-3' temperature of annealing (Ta) 56°C, *blaTEM-F* 5'-GCATACACTATTCTCAGAATGA-3' *blaTEM-R* 5'-CTCACCGGCTCCAGATTAT-3' Ta 56°C, *dfr7&17-F* 5'-CATTTGACTCTCTATGGGTGTTT-3' Ta 58°C.

To avoid analyzing duplicate resistant isolates, REP-PCR was performed on isolates showing the same phenotype and genotype, using the REP1 and REP2 primers as previously described¹⁴. *E. coli*-specific PCR was performed using primers to amplify *rrs* (75F and 619R) or *gad*^{15, 16}, with *E. coli* XL1-blue as a positive control.

Transformation and conjugation assays

Transferability of the detected antibiotic resistance genes was assessed by heat shock transformation using transformation competent *E. coli* DH5 α as the recipient strain and plasmid DNA obtained by alkaline lysis from all isolates resistant to tetracycline, ampicillin, chloramphenicol, or trimethoprim in which PCR had detected a resistance gene. Transformants were selected on LB agar plates containing either carbenicillin (50 μ g/mL), ampicillin (32 μ g/mL),

tetracycline (30 μ g/mL), chloramphenicol (30 μ g/mL), or trimethoprim (20 μ g/mL) as appropriate. Phenotypic resistance was confirmed for each transformant, and the presence of plasmid DNA and the relevant resistance genes were verified.

In some cases, the capacity for conjugation was assessed in liquid medium using *E. coli* J62-2 as the recipient strain. Selection was performed on LB agar plates supplemented with tetracycline (30 μ g/mL) and rifampicin (50 μ g/mL). Transconjugants were confirmed by phenotypic resistance, amplification of the same resistance gene detected in the donor, and ERIC-PCR with primers described by Versalovic *et al.*¹⁴.

RESULTS

Here, we characterized thirty-six isolates identified as *E. coli* with biochemical tests and identified as *E. coli* with biochemical tests and PCR amplification of *rrs* or *gad*. These isolates originated from the four sampling sites: 13 from Site 1, four from Site 2, nine from Site 3, and ten from Site 4.

Resistance phenotypes

As shown in Fig. 1, the highest frequency of resistance was to tetracycline and the lowest to ciprofloxacin and ampicillin-sulbactam, with intermediate prevalences of resistance to the other antibiotics tested, including ampicillin.

Twenty-four isolates (66.6%) were fully resistant to at least one antibiotic, corresponding to 7/13, 2/4, 7/9, and 8/10 isolates from sampling points 1 to 4, respectively (Table 1). Five of these 24 isolates (20.8%) also showed intermediate resistance to one or two additional antibiotics (two AMP-CIP, one CIP, and two SAM). Among the 12 remaining isolates, three had only intermediate resistance (two AMP and one TE), and nine (9/36, 25%) were fully sensitive to all antibiotics tested. There were also seven isolates (7/24, 29.1%) fully resistant to three or more antibiotics of different classes and

Table 1
Phenotypic and genotypic profiles of the bacterial isolates.

Sampling site	Resistance phenotype (FR/IR)	Resistance genotype	Phylogroup
1	TE	<i>tetB</i>	B1
1	TE-SF-WTS	<i>tetA-sul2</i>	C
1	TE-SF-WTS	<i>tetA-sul2</i>	C
1	TE-SF-WTS/amp-cip	<i>tetA-sul2</i>	C
1	TE-SF- TS-C/amp-cip	<i>tetA-sul2-floR</i>	A
1	TE-SF-C/ cip	<i>tetA-sul2-floR</i>	A
1	TE-C	<i>tetA-floR</i>	A
1	amp	nd	B1
1	S	na	C
1	S	na	C
1	S	na	C
1	S	na	C
1	S	na	B2
2	TE	<i>tetA</i>	C
2	TE-SF-WTS-AMP-C/sam	<i>tetA-sul1-blaTEM-floR-dfrA12-intI1</i>	C
2	S	na	A
2	S	na	A
3	TE	<i>tetA</i>	B1
3	TE-W	<i>tetA</i>	C
3	TE-SF-W- TS-AMP-C-CIP/sam	<i>tetA-sul3-blaTEM-intI1</i>	C
3	TE-SF-WTS- AMP-SAM	<i>tetA-sul3-blaTEM-dfrA12-intI1</i>	C
3	TE-AMP	<i>tetB-blaTEM</i>	C
3	AMP-C	<i>blaTEM- floR</i>	B1
3	AMP-W	<i>blaTEM</i>	C
3	te	nd	C
3	S	na	A
4	TE	<i>tetA</i>	C
4	TE	<i>tetA</i>	A
4	TE-SF	<i>tetA</i>	B1
4	TE-SF	<i>tetA-sul2</i>	A
4	TE-SF	<i>tetA-sul2</i>	C
4	TE-SF	<i>tetA-sul2</i>	B1
4	TE-SF-WTS-C	<i>tetA-sul3-intI1</i>	C
4	TE- W-C	<i>tetA- tetB- flor-intI1</i>	A
4	amp	nd	C
4	S	na	C

FR: Fully resistant, IR: intermediate resistance (lowercase), S: sensitive, nd: not determined, na: not apply. The abbreviations for the antibiotics are the same as in Fig. 1.

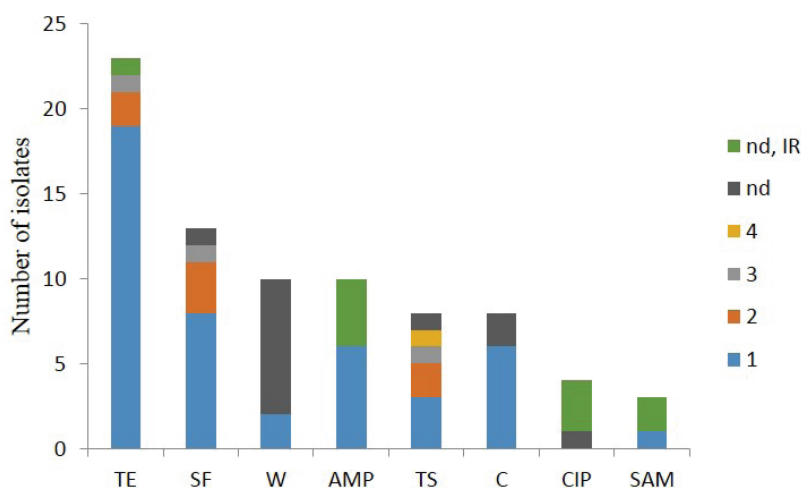


Fig. 1. Phenotypic and genotypic resistance per antibiotic. The numbers of phenotypically resistant isolates with corresponding genotypes (for full resistance) are indicated. Resistance genotypes are numbered from 1 to 4 according to the detection frequency from highest to lowest. 1: *tetA* (TE); *sul2* (SF and TS); *dfrA12* (W); *floR* (C); or *blaTEM* (AMP and SAM). 2: *tetB* (TE); or *sul3* (SF and TS). 3: *tetA-tetB* (TE); *sul1* (SF); *sul1-dfrA12* (TS). 4: *sul3-dfrA12* (TS). nd: Resistance genotype not determined, IR: intermediate resistance. TE, tetracycline; SF, Sulfisoxazole; W, trimethoprim; C, Chloramphenicol; TS, Trimethoprim-sulfamethoxazole; AMP, ampicillin; SAM, ampicillin-sulbactam; CIP, ciprofloxacin.

thus multi-drug resistant or MDR. Two of these originated from each of sampling Sites 1, 3, and 4, and one isolate was from Site 2 (see Table 1). All isolates were sensitive to aztreonam and imipenem.

Resistance genes

The genotypic resistance profiles and genes detected are described in Table 1. The most frequently detected genes were *tetA* (20/22) and *sul2* (8/12), while all isolates fully resistant to ampicillin contained *blaTEM* (Fig. 1). The *floR* gene was detected in most (6/8) of the chloramphenicol-resistant isolates, none of which contained the *cat* gene. The only amplified determinant associated with trimethoprim resistance was *dfrA12*, which was detected in just two of the ten trimethoprim-resistant isolates.

Of the seven MDR-resistant isolates, the *intI1* gene was amplified from 5, representing 21% of the resistant isolates and 13.9% of total isolates (Table 1). In one of the isolates in which the *intI1* gene was detected, ampli-

cation and sequencing with primers specific for conserved segments of class 1 integrons detected *dfrA12* and *aadA2*, which encodes an aminoglycoside adenylyl transferase conferring streptomycin resistance.

Transfer of resistance determinants

Transformants were obtained from the isolated plasmid DNA of 20/24 resistant isolates. Most of the transformants (18/20) were recovered on media with tetracycline, while only 1/20 were recovered on media with chloramphenicol and 1/20 on media with trimethoprim. However, the major part of the genes conferring resistance to sulfisoxazole and chloramphenicol were co-transferred with the *tetA* gene (Table 2). Despite repeated attempts, no transformants were obtained from plasmid DNA isolated from the remaining four resistant isolates.

The *tetA* gene was transferred from almost all isolates in which it was detected (19/20). The *blaTEM* gene was transferred

Table 2
Antibiotic resistance profiles of transformants and transconjugants.

Genotype	Phenotype	Number of transformants (N=20)	Number of transconjugants (N=2)
<i>tetA-sul2</i>	TE-SF	8	-
<i>tetA-sul3</i>	TE-SF	1	-
<i>tetA</i>	TE	6	2
<i>tetA-floR</i>	TE-C	3	-
<i>tetA- sul3- dfrA12</i>	TE-SF-W	1	-
<i>flor- blaTEM</i>	C-AMP	1	-

from 1/6 of the isolates in which it was detected, the *sul* genes were transferred from 10/12 isolates (8 *sul2* and 2 *sul3*), *floR* from 4/6 and *dfrA12* from 1/2 isolates containing this gene. The *tetB* gene was not transferred under the conditions employed. The resistance genes most frequently detected in our isolates, *tetA* and *sul2*, hybridized with plasmid DNA isolated from most (18/20), or all isolates (8/8), respectively, in which the genes had been detected by PCR, confirming that they were carried on plasmids (not shown).

The transfer capability of the *tetA* gene was tested by conjugation experiments with five isolates in which the *tetA* gene was detected. Transconjugants were obtained from two isolates at the high frequencies of 7.5×10^{-3} y 1.95×10^{-2} , demonstrating that the *tetA* gene was carried on conjugative plasmids at least in these two isolates. One of these five isolates with the *tetA* gene generated neither transformants nor transconjugants; nevertheless, only scant plasmid DNA could be obtained from this isolate, perhaps because its plasmid was either very large or present at a very low copy numbers.

***E. coli* phylogroup analysis**

Of the 36 bacterial isolates studied, 20 belonged to group phylogenetic C, nine to group A, six to group B1, and one to group B2 (Table 1).

DISCUSSION

In the present work, we analyzed phenotypic antibiotic-resistance, detected corresponding antibiotic-resistance genes, evaluated their transferability, and determined the phylogroups of 36 *E. coli* isolates obtained from a wastewater treatment plant. Similar to other studies of resistant *E. coli* isolates from wastewater and treatment plants¹⁷⁻²¹, the most frequent resistance encountered was to tetracycline, sulfonamide, trimethoprim, and ampicillin. In contrast, resistance to the carbapenems and ciprofloxacin was infrequent or not detected. Our results differ from a previous phenotypic study of *E. coli* isolates from stabilization ponds in Maracaibo, Venezuela, that found a higher proportion of resistance for ampicillin (81.25%) followed by tetracycline and trimethoprim²¹. In comparison, a study of *E. coli* isolates from a treatment plant in Cumaná (Venezuela) found 73.3% and 23.3% of ampicillin and tetracycline resistance, respectively. These discrepancies may be due to differences in the treatment systems.

Although tetracycline and sulfonamide, to which we found medium to high frequencies of resistance (22% to 64%), have been, for the most part, replaced by newer agents in human medicine, they are still classified as highly important by the World Health Organization.²³ Ampicillin (28% resistance) is

still frequently used and considered critically important in human medicine, as are the antibiotics for which we found less frequent resistance (8% to 11%).

The antibiotic resistance we observed is common in *E. coli* isolated from healthy humans in low and middle-income countries²⁴, and the antibiotic-resistance in human isolates of *E. coli* has been correlated with the resistance in *E. coli* isolated from the local wastewater²⁵. The resistance patterns may also be affected by selection or adaptation to the specific characteristics of the treatment plant²⁶. A high prevalence of the *tetA*, *sul2*, and *blaTEM* genes has been previously observed in *E. coli* isolated from other treatment plants.^{17, 27} However, previous studies have predominantly found the *cat* gene in chloramphenicol-resistant isolates^{17, 25}, but most of our chloramphenicol resistant isolates carried the *floR* gene. In contrast, the *cat* gene was not detected.

We found the *intI1* gene in 13.9% of total isolates, similar to a study by Figueira et al. that found the gene in 22.3 % of *E. coli* isolates from treated wastewater²⁰. Integrons contribute to the spread of multi-drug resistance, and class I integrons are the most important in clinical isolates²⁸.

Horizontal transmission of plasmids carrying resistance genes is a crucial route for disseminating resistance. Similar to our results, a previous study of *E. coli* and other fecal coliforms isolated from treatment plants found that most plasmid transformants were resistant to tetracycline, followed by chloramphenicol and trimethoprim²⁹. The resistance genes *tetA*, *blaTEM*, *tetB*, *sul*, *floR*, and *dfrA12* can be found on chromosomes, but as we observed, they are frequently carried on plasmids³⁰⁻³⁵.

E. coli isolates can be classified into seven main phylogenetic groups: A, B1, B2, C, D, E, and F. Intestinal pathogenic *E. coli* strains have been associated with groups A, B1, and E, while extra-intestinal strains mainly belong to groups B2 and D³⁶. However, Group C can also include human patho-

genic strains, as shown in a study of human isolates from the USA and Europe, in which phylogroup C was associated with uropathogenic *E. coli*, although B2 and D predominated³⁷. Jafari et al. found that group C was the most common (21.3%) among Shiga toxin-producing *E. coli* (STEC) patient isolates³⁸. It has been suggested that some *E. coli* phylogroups, particularly groups A and B1 (or non-B2), are more prone to develop antibiotic resistance to traditional antibiotics and fluoroquinolones^{39, 40}. Most previous studies assigning *E. coli* phylogroups have used the triple PCR method⁴¹ that identifies only four groups: A, B1, B2, and D. Several studies have found group A to predominate in wastewater, followed by D or B1²⁰. Group A is also the phylogroup most frequently observed in human commensal isolates, followed by B2 group⁴². Researchers employing the quadruplex method plus the group C specific PCR, or in silico typing, have found that phylogroup C is less frequent than other phylogroups in samples from birds, humans, non-human mammals, domestic animals, wild animals, river and lake water⁴³⁻⁴⁵. Two studies on strains from wastewater found that group A or B2 predominated, while none or only 1% belonged to group C^{17, 46}.

Therefore, the presence of group C in most of our isolates differs from the findings in similar studies and could be due to geographical location or climate differences, as these factors may influence the distribution of phylogroups⁴⁵. It is also possible that previous studies that used only the triple PCR method classified C group isolates as group A.

In conclusion, we observed *E. coli* isolates resistant to diverse antibiotics, including some clinically essential agents and found that many were associated with transferable genetic determinants and class 1 integrons. Also, in contrast to other studies, many of our isolates belonged to the phylogroup C. The characteristics of the isolates we studied may have been determined by the influent water, the nature of the treatment plant, and environmental conditions, but an evalu-

ation of the contribution of each of these aspects would require a much larger study with many more isolates. Similar studies should be repeated in the same treatment plant and undertaken in other treatment plants in Venezuela, and these studies should be extended to include untreated wastewater and focus on resistance to the newer, currently more commonly used antibiotics.

Conflict of interest

The authors declare that they have no conflict of interest.

Funding

The Instituto Venezolano de Investigaciones Científicas supported this research.

ORCID numbers

- Elba Guerrero (EG):
0000-0002-3936-9556
- Lizeth Caraballo (LC):
0000-0003-2043-8731
- Howard Takiff (HT):
0000-0002-0480-0860
- Dana García (DG):
0009-0007-7020-5529
- Marynes Montiel (MM):
0000-0002-6249-0362

Author contributions

Methodology: EG, LC, DG, and MM. Data analysis and original draft preparation: EG. Supervision, writing – review, and editing: HT. All authors contributed to the study conception, commented on previous versions, and read and approved the final manuscript.

REFERENCES

1. Anjum MF, Schmitt H, Börjesson S, Berendonk TU; WAWES network. The potential of using *E. coli* as an indicator for the surveillance of antimicrobial resistance (AMR) in the environment. *Curr Opin Microbiol* 2021; 64:152-158.
2. Grehs BWN, Linton MAO, Clasen B, de Oliveira Silveira A, Carissimi E. Antibiotic resistance in wastewater treatment plants: understanding the problem and future perspectives. *Arch Microbiol* 2021; 203(3): 1009-1020.
3. American Public Health Association. Standard Methods For the Examination of Water and Wastewater. 20 th Ed. Washington DC: APHA; 2005, p 2215-2218.
4. Clinical and Laboratory Standards Institute. Performance Standards for Antimicrobial Susceptibility testing. Twenty-Fifth Informational Supplement M100-S25. Wayne PA, USA: CLSI; 2015, p 44-50.
5. Jones CH, Tuckman M, Murphy E, Bradford PA. Identification and sequence of a *tet(M)* tetracycline resistance determinant homologue in clinical isolates of *Escherichia coli*. *J Bacteriol* 2006; 188(20):7151-7164.
6. Van TT, Chin J, Chapman T, Tran LT, Coloe PJ. Safety of raw meat and shellfish in Vietnam: an analysis of *Escherichia coli* isolations for antibiotic resistance and virulence genes. *Int J Food Microbiol* 2008; 124(3):217-223.
7. Lanz R, Kuhnert P, Boerlin P. Antimicrobial resistance and resistance gene determinants in clinical *Escherichia coli* from different animal species in Switzerland. *Vet Microbiol* 2003; 91(1):73-84.
8. Perreten V, Boerlin P. A new sulfonamide resistance gene (*sul3*) in *Escherichia coli* is widespread in the pig population of Switzerland. *Antimicrob Agents Chemother* 2003; 47(3):1169-1172.
9. Fang H, Ataker F, Hedin G, Dornbusch K. Molecular epidemiology of extended-spectrum beta-lactamases among *Escherichia coli* isolates collected in a Swedish hospital and its associated health care facilities from 2001 to 2006. *J Clin Microbiol* 2008; 46(2):707-712.
10. Guerra B, Soto SM, Argüelles JM, Mendoza MC. Multidrug resistance is mediated

- by large plasmids carrying a class 1 integron in the emergent *Salmonella enterica* serotype [4,5,12:i:-]. Antimicrob Agents Chemother 2001; 45(4):1305-1308.
11. Grape M, Motakefi A, Pavuluri S, Kahlmeter G. Standard and real-time multiplex PCR methods for detection of trimethoprim resistance *dhfr* genes in large collections of bacteria. Clin Microbiol Infect 2007; 13(11):1112-1118.
 12. Moura A, Henriques I, Ribeiro R, Correia A. Prevalence and characterization of integrons from bacteria isolated from a slaughterhouse wastewater treatment plant. J Antimicrob Chemother 2007; 60(6):1243-1250.
 13. Clermont O, Christenson JK, Denamur E, Gordon DM. The Clermont *Escherichia coli* phylo-typing method revisited: improvement of specificity and detection of new phylo-groups. Environ Microbiol Rep 2013; 5(1):58-65.
 14. Versalovic J, Koeuth T, and Lupski JR. Distribution of repetitive DNA sequences in eubacteria and application to fingerprinting of bacterial genomes. Nucleic Acids Res 1991; 19(24): 6823-6831.
 15. Sabat G, Rose P, Hickey W.J and Harkin M. Selective and sensitive methods for PCR amplification of *Escherichia coli* 16S rRNA genes in Soil. Appl Environ Microbiol 2000; 66 (2): 844-849.
 16. McDaniels AE, Rice EW, Reyes AL, Johnson CH, Haugland RA, Stelma GN Jr. Confirmational identification of *Escherichia coli*, a comparison of genotypic and phenotypic assays for glutamate decarboxylase and beta-D-glucuronidase. Appl Environ Microbiol 1996; 62(9):3350-3354.
 17. Ogura Y, Ueda T, Nukazawa K, Hiroki H, Xie H, Arimizu Y. The level of antimicrobial resistance of sewage isolates is higher than that of river isolates in different *Escherichia coli* lineages. Sci Rep 2020;10(1):17880.
 18. Paulshus E, Kuhn I, Mollby R, Colque P, O'Sullivan K, Midtvedt T, Lingaas E, Holmstad R, Sørum H. Diversity and antibiotic resistance among *Escherichia coli* populations in hospital and community wastewater compared to wastewater at the receiving urban treatment plant. Water Res 2019; 161:232-241.
 19. Blaak H, Lynch G, Italiaander R, Hamidjaja RA, Schets FM, de Roda Husman AM. Multidrug-resistant and extended spectrum beta-lactamase-producing *Escherichia coli* in Dutch surface water and wastewater. PLoS One 2015; 10(6):e0127752.
 20. Figueira V, Serra E, Manaia CM. Differential patterns of antimicrobial resistance in population subsets of *Escherichia coli* isolated from waste- and surface waters. Sci Total Environ 2011; 409(6):1017-1023.
 21. Zambrano J L, Botero L, Cavazza M E, Avila M. Resistencia a antimicrobianos y presencia de plásmidos en cepas de *Escherichia coli* aisladas de aguas residuales crudas y tratadas por lagunas de estabilización con fines de reuso en agricultura. Rev Soc Ven Microbiol 2002; 22(1): 44-50
 22. Martínez, RE, Villalobos LB. Susceptibilidad antimicrobiana de cepas de *Escherichia coli* aisladas de alimentos y aguas residuales en Cumaná, Venezuela. SABER 2008; 20(2), 172-176.
 23. World Health Organization. Critically important antimicrobials for human medicine. 6th revision. Geneva: WHO; 2019, p 14-15.
 24. Nji E, Kazibwe J, Hambridge T, Joko CA, Larbi AA, Dampney LAO, Nkansa-Gyamfi NA, Stålsby Lundborg C, Lien TQ. High prevalence of antibiotic resistance in commensal *Escherichia coli* from healthy human sources in community settings. Sci Rep 2021; 11(1):3372.
 25. Raven KE, Ludden C, Gouliouris T, Blane B, Naydenova P, Brown NM, Parkhill J, Peacock SJ. Genomic surveillance of *Escherichia coli* in municipal wastewater treatment plants as an indicator of clinically relevant pathogens and their resistance genes. Microb Genom 2019; 5(5): e000267.
 26. Turolla A, Cattaneo M, Marazzi F, Mezzanotte V, Antonelli M. Antibiotic resistant bacteria in urban sewage: Role of full-scale wastewater treatment plants on environmental spreading. Chemosphere 2018;191:761-769.

27. Osinska A, Korzeniewska E, Harnisz M, Niestepski S. The prevalence and characterization of antibiotic-resistant and virulent *Escherichia coli* strains in the municipal wastewater system and their environmental fate. *Sci Total Environ* 2017; 577:367-375.
28. Souque C, Escudero JA, MacLean RC. Integron activity accelerates the evolution of antibiotic resistance. *Elife* 2021; 10: e62474.
29. Smyth C, O'Flaherty A, Walsh F, Do TT. Antibiotic resistant and extended-spectrum β -lactamase producing faecal coliforms in wastewater treatment plant effluent. *Environ Pollut* 2020; 262:114244.
30. Wang Y, Batra A, Schulenburg H, Dagan T. Gene sharing among plasmids and chromosomes reveals barriers for antibiotic resistance gene transfer. *Philos Trans R Soc Lond B Biol Sci* 2022; 377(1842):20200467.
31. Hansen KH, Andreasen MR, Pedersen MS, Westh H, Jelsbak L, Schonning K. Resistance to piperacillin/tazobactam in *Escherichia coli* resulting from extensive IS26-associated gene amplification of *bla-TEM-1*. *J Antimicrob Chemother* 2019; 74(11):3179-3183.
32. Sunde M, Sorum H. Self-transmissible multidrug resistance plasmids in *Escherichia coli* of the normal intestinal flora of healthy swine. *Microb Drug Resist* 2001; 7(2):191-196.
33. Jiang H, Cheng H, Liang Y, Yu S, Yu T, Fang J, Zhu C. Diverse mobile genetic elements and conjugal transferability of sulfonamide resistance genes (*sul1*, *sul2*, and *sul3*) in *Escherichia coli* isolates from penaeus vannamei and pork from large markets in Zhejiang, China. *Front Microbiol* 2019;10:1787.
34. Meunier D, Jouy E, Lazizzera C, Doublet B, Kobisch M, Cloeckaert A, Madec JY. Plasmid-borne florfenicol and ceftiofur resistance encoded by the *floR* and *blaCMY-2* genes in *Escherichia coli* isolates from diseased cattle in France. *J Med Microbiol* 2010; 59(Pt 4):467-471.
35. Yu HS, Lee JC, Kang HY, Jeong YS, Lee EY, Choi CH, Tae SH, Lee YC, Seol SY, Cho DT. Prevalence of *dfr* genes associated with integrons and dissemination of *dfrA17* among urinary isolates of *Escherichia coli* in Korea. *J Antimicrob Chemother* 2004; 53(3):445-450.
36. Clermont O, Olier M, Hoede C, Diancourt L, Brisse S, Keroudean M, Glodt J, Picard B, Oswald E, Denamur E. Animal and human pathogenic *Escherichia coli* strains share common genetic backgrounds. *Infect Genet Evol* 2011; 11(3):654-662.
37. Biggél M, Xavier BB, Johnson JR, Nielsen KL, Frimodt-Møller N, Matheeussen V, Goossens H, Moons P, Van Puyvelde S. Horizontally acquired papGII-containing pathogenicity islands underlie the emergence of invasive uropathogenic *Escherichia coli* lineages. *Nat Commun* 2020; 11(1):5968.
38. Jafari E, Oloomi M, Bouzari S. Characterization of antimicrobial susceptibility, extended-spectrum β -lactamase genes and phylogenetic groups of Shigatoxin producing *Escherichia coli* isolated from patients with diarrhea in Iran. *Ann Clin Microbiol Antimicrob* 2021; 20(1), 24.
39. Citterio B, Andreoni F, Simoni S, Carloni E, Magnani M, Mangiaterra G, Cedrarò N, Biavasco F, Vignaroli C. Plasmid replicon typing of antibiotic-resistant *Escherichia coli* from clams and marine sediments. *Front Microbiol* 2020; 11:1101.
40. Sabaté M, Prats G, Moreno E, Ballesté E, Blanch AR, Andreu A. Virulence and antimicrobial resistance profiles among *Escherichia coli* strains isolated from human and animal wastewater. *Res Microbiol* 2008; 159(4):288-93.
41. Clermont O, Bonacorsi S, Bingen E. Rapid and simple determination of the *Escherichia coli* phylogenetic group. *Appl Environ Microbiol* 2000; 66(10):4555-4558.
42. Tenaillon O, Skurnik D, Picard B, Denamur E. The population genetics of commensal *Escherichia coli*. *Nat Rev Microbiol* 2010; 8(3):207-217.

43. Lescat M, Clermont O, Woerther PL, Glodt J, Dion S, Skurnik D, Djossou F, Dupont C, Perroz G, Picard B, Catzefflis F, Andremont A, Denamur E. Commensal *Escherichia coli* strains in Guiana reveal a high genetic diversity with host-dependant population structure. *Environ Microbiol Rep* 2013; 5(1):49-57.
44. Petit F, Clermont O, Delannoy S, Servais P, Gourmelon M, Fach P, Oberlé K, Fournier M, Denamur E, Berthe T. Change in the structure of *Escherichia coli* population and the pattern of virulence genes along a rural aquatic continuum. *Front Microbiol* 2017; 8:609.
45. Touchon M, Perrin A, de Sousa JAM, Vangchhia B, Burn S, O'Brien CL, Denamur E, Gordon D, Rocha EP. Phylogenetic background and habitat drive the genetic diversification of *Escherichia coli*. *PLoS Genet* 2020 ;16(6):e1008866.
46. Zhi S, Banting G, Li Q, Edge TA, Topp E, Sokurenko M, Scott C, Braithwaite S, Ruecker NJ, Yasui Y, McAllister T, Chui L, Neumann NF. Evidence of naturalized stress-tolerant strains of *Escherichia coli* in Municipal wastewater treatment plants. *Appl Environ Microbiol* 2016; 82(18):5505-5518.